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36

37 **Introduction**

38 Wastewater analysis (WWA) or Sewage epidemiology was first proposed to estimate
39 drug consumption by US EPA environmental scientist Daughton in 2001 (Daughton,
40 2001). The approach was based on the assumption that when particular drugs are
41 consumed, the active parent compounds and its metabolites are excreted through urine
42 and faeces into the sewer system, and thus enter the sewage treatment plants (STPs).
43 By measuring levels of target parent compounds and/or metabolites, back-estimation
44 of drug use in the population of a STP catchment area could be realised. Compared
45 with conventional methods such as questionnaires and socio-epidemiological surveys
46 including crime statistics, medical records, drug production and seizure rates, WWA
47 has the advantage of providing objective, continuous, near real-time estimates of drug
48 consumption in the population (van Nuijs et al., 2011a). Additionally, using WWA to
49 estimate illicit drug consumption can overcome ethical issues associated with some
50 other methods (Hall et al., 2012; Khan et al., 2014).

51 A lot of effort has been made to improve all aspects of WWA. These include
52 sampling protocol development to get representative samples, developing robust,
53 sensitive analytical methods and more recently normalizing chemical loads to per
54 capita estimates that allows more accurate comparisons between different cities and
55 even countries (Ort et al., 2010a; Ort et al., 2010c; Zuccato et al., 2011; Zuccato et al.,
56 2005, O'Brien et al., 2014). Many researchers from a wide range of fields including
57 but not limited to analytical chemistry, environmental science, epidemiology, forensic
58 science, sociology and statistics from all over the world have joined the 'WWA
59 research community' to improve the innovative approach during the past years. This
60 is evident by a series of conferences organised by the European Monitoring Centre for
61 Drug and Drug Addiction called *Testing the waters* starting in May 2013 in Lisbon,
62 Portugal and the next session is planned for 2015 in Ascona, Switzerland.

63 This review article attempted to present a brief overview of the development of WWA
64 to date with a focus on its successful application to estimate illicit drug consumption
65 and the future applicability of this approach in China.

66 **1. Current state of WWA**

67 **1.1 Application of WWA in estimating illicit drug consumption**

68 The approach of WWA was applied for the first time in Italy in 2005 (Zuccato et al.,
69 2005) and was soon applied in several other cities in Europe and the US (van Nuijs et
70 al., 2011a). Since then WWA has been applied to monitor the use of the classical
71 illicit drugs such as cocaine, heroin, amphetamines and cannabis (Thomas et al., 2012;
72 van Nuijs et al., 2011a) and more recently to identify the use of new psychoactive
73 substances (Reid et al., 2014; van Nuijs et al., 2014).

74 Reports of illicit drugs estimated by WWA have come from multiple countries
75 including Australia, Belgium, Canada, Croatia, Finland, France, Italy, Ireland, The
76 Netherlands, Sweden, the UK and the US. Estimation of illicit drug use has been
77 performed not only in small communities such as prisons (Postigo et al., 2011), and
78 recreational regions (Lai et al., 2012), but also in large cities like Paris and Hong
79 Kong (Karolak et al., 2010; Lai et al., 2013). Most obtained results are in agreement
80 with data from traditional socio-epidemiological surveys, however some
81 underestimation and/or overestimation has been identified for some particular drug(s)
82 (Baker et al., 2014). Thomas et al. (2012) conducted a comparison of illicit drug
83 consumptions in 19 cities across Europe through WWA and identified distinct
84 temporal and spatial differences in drug consumption between these cities during a
85 single week of sampling in 2011. In 2013, Nefau et al. (2013) studied the presence of
86 17 illicit drugs both in influent and effluent sewage from 25 French STPs.
87 Consumption maps were drawn for cocaine, opiates, cannabis and amphetamine-like
88 compounds. Significant geographical differences were observed which highlighted
89 that drug consumption within a country might not be homogeneous. Similarly, Khan
90 et al (2014) applied WWA to evaluate the use of 10 illicit drugs in 4 megacities in
91 China and found different consumption patterns between north and south China. At
92 the same time, Li et al. (2014) also reported the use of amphetamines across a range
93 of communities in the metropolitan area of Beijing. A summary of WWA applications
94 for assessing illicit drug consumption worldwide is shown in **Table 1**.

96 **1.2 Exploration in other areas**

97 In addition to illicit drugs, there are some initial attempts to estimate the use of
98 alcohol and tobacco, the two most common substances that have potential to
99 negatively impact population health and cause several social problems such as crime
100 and injuries. Reid et al conducted the first study to estimate the use of alcohol in Oslo,
101 Norway using WWA (Reid et al., 2011) where the highest consumption of the alcohol
102 was observed during weekends. Sixty one percent of weekly alcohol consumption was
103 reported on Friday and Saturday alone. Over the last year, two studies were carried
104 out to estimate the total amount of tobacco use (nicotine consumption) in different
105 communities through WWA (Castiglioni et al., 2014; Lopes et al., 2014). The
106 findings produced by WWA were in close agreement with survey data and can
107 differentiate the level of tobacco consumption among different populations.

108 In addition to monitoring common substances of abuse (illicit drugs, alcohol, and
109 tobacco), WWA can be considered as Sewage chemical-information mining (SCIM).
110 In a broader sense, because the interpretation of acquired information from WWA can
111 measure a vast amount of chemicals, WWA can provide a variety of information
112 about the population living in a particular STP catchment. It could also be used as a
113 powerful tool to evaluate community-wide human health with isoprostanes (stress
114 biomarkers) already proposed by Daughton as ideal candidates (Daughton, 2012b).
115 Daughton also conceptualized an approach to estimate the real-time population size in
116 the sewer catchment using coprostanol as a population biomarker (Daughton, 2012a)
117 although further study should be conducted to validate the applicability of coprostanol
118 in WWA (Chen et al., 2014).

119 Venkatesan and Halden have applied SCIM to forecast ecological and human health
120 risks of manmade chemicals by analysing sewage sludge instead of wastewater for
121 persistent organic pollutants (POPs) which are non-polar and thus less likely to be in
122 the wastewater itself (Venkatesan and Halden, 2014). The result revealed 70%
123 agreement between WWA and biological specimens' analysis, and suggested that
124 analysing sewage sludge can inform human health risk assessments by providing real-
125 time information on toxic exposures in human populations and associated body
126 burdens of harmful, accumulative, environmental pollutants. More outcomes could be
127 achieved if the efforts across several disciplines including clinical chemistry,

128 environmental chemistry, environmental science, medicine and microbiology, and
129 were combined. With continuous improvement of the method, SCIM appears a
130 feasible and effective tool to identify the connection between population health and
131 chemical consumption and/or exposure and thus enabling better protection of the
132 population from such hazards.

133 **1.3 Current research to improve the methodology**

134 Current research mostly focuses on evaluating and minimizing uncertainties of the
135 whole WWA procedure such as collecting representative sewage samples, simplifying
136 sample pre-treatment, selecting suitable biomarkers in terms of sensitivity and
137 stability, optimizing instrumental analysis, and refining the back calculation of results.
138 Castiglioni et al. integrally addressed uncertainties associated with all the steps
139 necessary to estimate community drug consumption through WWA (Castiglioni et al.,
140 2013). Using data gathered from 12 laboratories, the uncertainties can range from 5–
141 10% for sampling to 1–34% for replicated chemical analysis and 26% for back-
142 calculation of cocaine use. But the highest uncertainty comes from the estimation of
143 population size, which varied from 7 to 55%. Based on this study, the authors also
144 suggested a best practice protocol to minimize the overall uncertainties of the entire
145 procedure (Castiglioni et al., 2013).

146 Several studies have attempted to address individual issues facing WWA. For instance,
147 Martínez Bueno et al. developed a solvent-free method for simultaneous identification
148 and quantification of 22 illicit drugs by liquid chromatography coupled to tandem
149 mass spectrometry (LC-MS/MS), which is deemed to be a good technique for WWA
150 due to its simplicity, cost-effectiveness and lower environmental footprint (Martínez
151 Bueno et al., 2011). Meanwhile, Baker and Kasprzyk-Hordern evaluated the
152 commonly used methodologies for sample collection, storage and preparation for
153 SCIM with solid-phase extraction (SPE) and LC-MS/MS analysis (Baker and
154 Kasprzyk-Hordern, 2011). They concluded that from the perspective of stability,
155 composite samples are unsuitable with regards to certain compounds like heroin and
156 6-acetylmorphine; these two drugs reported a decrease in stability of 66% and 26%
157 respectively after 12 hours in raw sewage at 2°C. Baker and Kasprzyk-Hordern also
158 emphasised that more rigorous reporting of method validation data are needed as
159 underreported parameters might have major impacts on the overall performance.

160 For the estimation of consumed drug masses in the catchment using the optimum
161 sampling method as outlined by Ort et al. (2010b) and common chemical analysis, Lai
162 et al. calculated the overall uncertainty to be in the range of 20-30% (relative standard
163 deviation, RSD) (Lai et al., 2011). Lai et al. also suggested using chemicals of
164 relatively high use in the population as a basis to estimate the population size. To
165 further address this issue, O'Brien et al. have screened wastewater samples and found
166 14 chemicals which could be use as real-time population markers. They then
167 developed a model to estimate the population contributing to the sewage influents
168 based on the load of those chemicals. Through calibrating their model with mass loads
169 of 14 chemicals with accurate population counts (the samples were taken on Census
170 day), they found that relatively accurate population sizes can be estimated for
171 catchment >100,000 people (O'Brien et al., 2014).

172 **2 General procedure of WWA**

173 WWA is generally carried out using the procedure shown below (**Fig. 1**).
174 Simplification and standardization of the method as well as improvement of the
175 accuracy and reliability of the final estimates are crucial in promoting WWA for
176 routine monitoring.

177 **2.1 Pre-investigation**

178 A systematic and comprehensive pre-investigation about the catchment area and STP
179 under investigation is critical for reliable and accurate WWA estimates. Socio-
180 economic conditions of the study area, contemporary and historical environmental
181 monitoring data, population size and mobility in and out of the STP catchment area,
182 and crime statistics should all be put into consideration to achieve reliable results. The
183 investigation could be carried out through multiple approaches such as literature
184 reviews, visiting and surveying STPs, discussions with local authorities such as law
185 enforcement officers, relevant medical staff as well as environmental officers, local
186 and national councillors etc. The pre-investigation may strengthen the results'
187 reliability of WWA studies particularly where drug consumption estimates are the
188 goal. These alternate methods for assessing community drug consumption are not
189 limited to the pre-investigation stage but are also may be relevant to reconsider during
190 or even post the sampling period. Examples of this include combining drug seizure

191 data with loads in the wastewater and assessing the scale of the market based on the
192 mass load of drugs removed.

193 **2.2 Sampling**

194 Samples are taken from the inlets of STPs since the influent can be regarded as a
195 pooled urine sample (although diluted and contaminated) from a large population
196 before it is altered by different treatment processes in the STPs. However, in addition
197 to sewage influent, activated sludge from the aerobic or anaerobic tanks has also been
198 used as samples for WWA (Venkatesan and Halden, 2014).

199 For sample volume, one litre is the most common. However, sample volumes from
200 0.05 to 10 litres have been reported. A variety of sampling methods have also been
201 studied. Continuous flow, volume and time proportional sampling with
202 commercialized auto-samplers have all been used in different studies. Grab samples
203 have also been used in several studies (Hummel D, 2006). Ort et al. (2010b) found
204 continuous flow proportional samples collected over a 24 hour period as the optimum
205 sampling method as these are more representative of a whole day and are better at
206 capturing events. Samples from weekdays, weekends and public holidays across the
207 whole year have all been investigated to reveal temporal patterns of drug consumption.
208 While it is possible for each research group to establish a continuous flow
209 proportional sampling system, there is a need for the development of a commercial
210 auto-sampler that use this optimal sampling method. This would allow for a
211 standardized sampling approach for WWA while sampling at the different STPs (Ort
212 et al., 2010b).

213 Detailed discussion about sampling practices for wastewater has been conducted and
214 a comprehensive sampling guide with the aim of reducing uncertainties has been
215 proposed (Ort et al., 2010c). Evaluation of flow measurement, choice of sampling
216 mode, determination of frequency and location have all been discussed in the
217 abovementioned paper. For long-term routine monitoring, on-line auto-samplers are
218 essential for representative sampling with reliability, efficiency and from economic
219 aspects. More research should be conducted to evaluate uncertainties brought by
220 different sampling parameters in the future.

221 **2.3 Biomarker selection**

222 Selection of suitable biomarkers is an important factor for WWA. There are several
223 criteria for appropriate WWA biomarkers as suggested by Daughton (Daughton,
224 2012b) including: produced exclusively by humans (i.e., not introduced by unrelated,
225 exogenous mechanisms, e.g. illicit drug discharge), excreted in sufficient quantities
226 (to allow detection in sewage), sufficiently stable in the sewer pipeline, amenable to
227 cost-effective, reproducible analysis, and for several health status biomarkers they
228 should be excreted at elevated levels under “stressed condition” significantly different
229 to the baseline range of the chemicals excreted under “normal condition” .

230 While biomarkers have been one of the most popular research topics in clinical
231 science in the past decades, there were limited studies on biomarkers that can be used
232 in WWA. As suggested by Daughton, one should start at the list of common clinical
233 biomarkers and test them against the appropriate criteria (Daughton, 2012b). One of
234 the criteria that has been tested in several studies is the stability of the biomarkers in
235 the wastewater matrix and under sewer conditions. Until recently, most parent
236 compounds and metabolites were used as biomarkers in WWA for monitoring of
237 illicit drug consumption with the assumption that they were stable in the sewer system.
238 But some of these biomarkers (such as cocaine or 6 acetyl morphine) are quite
239 unstable (Thai et al., 2014; van Nuijs et al., 2012) which means that previous studies
240 may have underestimated the amount of drugs consumed in certain catchments. To
241 address this, excretion profiles of biomarkers including parent to metabolite ratio
242 should be further investigated by pharmacologists, biochemists and sewer engineers
243 to get a better grasp of consumed load versus measured load within wastewater.

244 For WWA to reach its’ full potential, more biomarkers should be identified and tested
245 against all of Daughton’s proposed criteria to expand the WWA application to
246 evaluate other markers of population health, real-time population size estimation,
247 pollutant exposure, and promote WWA as a routine monitoring approach in STPs.

248 **2.4 Pre-treatment**

249 Filtration or centrifugation of the collected sample is essential to remove solids in the
250 sample. However, this step may cause loss of certain analytes due to substantial
251 affinity for particulate for some chemicals (Baker and Kasprzyk-Hordern, 2011;

252 Plo'sz et al., 2013). Adding isotope labelled internal standards before filtration or
253 centrifugation is an effective approach to evaluate and minimize these uncertainties.
254 Full and accurate understanding about biomarker absorption kinetics is also useful to
255 minimize the uncertainties associated with correction factors for the back calculation
256 process.

257 The observed concentrations of target compounds and their metabolites in raw sewage
258 are often at the level of ng/L or even lower and thus pre-concentration is required. In
259 most cases solid-phase extraction (SPE) is conducted prior to LC-MS/MS analysis in
260 order to concentrate and remove matrix interferences from the samples.

261 Baker and Kasprzyk-Hordern have critically evaluated the whole sample preparation
262 process from sample collection to storage and preparation for analysis. This was
263 conducted for both pharmaceuticals and illicit drugs in surface water and wastewater
264 using SPE-LC-MS/MS techniques (Baker and Kasprzyk-Hordern, 2011). The study
265 showed that uncertainties associated with biomarker degradation can be minimized if
266 proper pre-treatment is applied. The current optimal method is to collect samples in a
267 refrigerated (4°C) container, subsample them, acidify with hydrochloric acid and then
268 either refrigerate at 4°C in the dark or freeze if the samples are to be analysed at a later
269 date to minimize biotransformation/degradation of the biomarkers. Degradation of
270 illicit drugs and metabolites in wastewater has been evaluated by van Nuijs et al
271 (2012). They concluded that most parent compounds and metabolites of illicit drugs
272 such as amphetamine, methamphetamine, ecstasy and EDDP are considerably stable
273 for 12 hours or longer, however some drugs such as cocaine and ecgonine methylester
274 showed a clear decrease in concentration over this period.

275 Since the SPE process is costly, time consuming and requires larger sample volumes,
276 simpler procedures are starting to be developed. Berset et al developed a large volume
277 direct injection method for the simultaneous analysis of licit and illicit drugs in
278 surface water and waste water (Berset et al., 2010). It should be noted that analytical
279 instruments are becoming more sensitive and when combined with the development
280 of optimised methods, it seems plausible that reliable methods for analysing illicit
281 drugs in wastewater with acceptable limits of detection (LOD) without the need for
282 SPE is possible. This would then enhance the argument for routine WWA monitoring
283 as a tool for measuring drug consumption. The improved sensitivity of some
284 instruments (i.e. LC-MS/MS) is already adequate for determination of numerous

285 chemicals in wastewater using simple pre-treatment technique such as acidifying and
286 filtering only (e.g. the pharmaceuticals in O'Brien et al., 2014).

287 **2.5 Instrumental Analysis**

288 Liquid chromatography coupled with tandem mass spectrometry (LC-MS/MS) is used
289 in almost all WWA studies due to its high sensitivity, versatility and selectivity. A
290 variety of mass spectrometers have been used in different WWA studies (Castiglioni
291 et al., 2011). These include triple quadrupole mass spectrometer (QqQ), Orbitrap, and
292 quadrupole time-of-flight mass spectrometers (QTOF). Most applications use QqQ
293 since it provides better selectivity and thus can achieve low detection limits. For some
294 compounds the use of QqQ may eliminate the need for sample extraction and clean-
295 up by using direct-injection technique (Trenholm and Snyder, 2011). Since QTOF and
296 linear ion trap fourier transform (LIT-FT) have higher mass resolution and mass
297 accuracy than QqQ, they can be better choices for drug identification and new
298 synthetic drug screening in samples with complex matrix (de Voogt et al., 2011;
299 Hernández et al., 2011).

300 Multiple reaction monitoring (MRM) mode is used for both qualification and
301 quantification. Several studies grouped illicit drugs according to their
302 physicochemical properties and different analytical parameters (such as
303 chromatography and ionization mode) in order to achieve optimum separation and the
304 highest MS resolution. As newer, more sensitive LC-MS/MS, QTOF and Orbitrap
305 instruments are developed, it is expected that these will have a significant role in
306 WWA applications for their incomparable advantages in rapid wide-scope screening
307 and providing accurate mass data of both parent molecules and daughter ions for
308 identification in complex matrices. Some of these instruments such as QTOF and
309 Orbitrap have the ability to acquire both qualitative and quantitative information from
310 samples in one injection and thus chemicals of interest can be retrospectively identify
311 in these samples. Once the spectra are obtained, there is no need for reanalysing the
312 sample. This is particularly useful for emerging drugs or pollutants where the
313 analytical chemists usually have to catch-up to those producing the chemicals. In the
314 future, standardised configurations could be implemented for routine illicit drug
315 consumption monitoring while customised configuration will play an important role
316 in the expanding WWA applications.

317 **2.6 Back estimation of consumption/exposure data**

318 The estimation of illicit drug consumption (IDC) in the population is carried out by
319 using the equation below (Zuccato et al., 2008):

$$320 \text{ IDC (mg/person/day)} = \frac{C_i * F * \frac{R_i}{E_i}}{P}$$

321

322 Where C_i is the concentration of a given drug residue i (parent drug or metabolite)
323 measured in raw sewage samples (mg/L), F is the total flow during the sampling
324 period (L , 24 hours), P is the number of people in the catchment, R_i is the ratio of
325 molar mass of parent drug to its metabolite and E_i is the average excretion rate of a
326 drug residue i .

327 While C_i , F and R_i can be measured readily in the laboratory or at the STP, estimating
328 the values of P and E_i is more challenging. E_i can be estimated through meta-analysis
329 of clinical data (Khan and Nicell, 2011). Meanwhile, estimation of population size
330 could be performed by using resources like census data, STP design capacity, or using
331 wastewater parameters such as BOD, COD, total phosphorus, total nitrogen (van
332 Nuijs et al., 2011b). O'Brien et al used a combination of 14 chemical markers of
333 population size (most of them pharmaceuticals) to estimate the population size using a
334 Bayesian inference model (O'Brien et al., 2014). Chen et al evaluated seven potential
335 population biomarkers and found that 5-hydroxyindoleacetic acid and cotinine could
336 potentially be used as biomarkers for population estimation (Chen et al., 2014). There
337 are also attempts to evaluate real-time population size by analysing mobile phone
338 signals in the catchment area which could also be applied for population estimation
339 (Ran et al., 2013).

340 It should be noted that there may be some licit sources of biomarkers used to estimate
341 illicit drugs (e.g. morphine can be generated from the consumption of both heroin and
342 licit codeine) and hence estimates of illicit drug consumption can be affected by this
343 phenomenon. In such cases, cautious interpretation should be taken. The typical way
344 to solve this issue is to subtract the average amount of legal medication/pharmaceuticals
345 that are used in the studied population from the total chemical load measured in
346 wastewater samples. The input load coming from licit source could be better
347 evaluated by analysing prescription data and wastewater from the hospitals in the

348 studied catchment. If the input from licit source is significant (e.g. morphine input
349 from the use of codeine compared to morphine input from heroin), the lack of
350 accurate data on licit input could render the WWA estimate less valid and thus WWA
351 should not be used in such case.

352 For other chemicals, the process of back estimation is similar as long as the necessary
353 parameters are available (especially E_i and P). Some chemicals may also come from
354 other sources such as dumping parent compounds into the sewer which should be
355 taken into account when interpreting the estimated values.

356 **2.7 Uncertainties and Limitations**

357 Uncertainties may occur in every step from sampling to back-calculation in WWA
358 studies. Evaluations about uncertainties related to the whole procedure and also
359 individual aspects have been performed in previous studies. Castiglioni et al
360 (Castiglioni et al., 2013) evaluated uncertainties associated with all the steps
361 commonly used in WWA with optimized experimental parameters for each step
362 defined. Plósz et al. (2013) investigated the biotransformation kinetics and sorption of
363 cocaine and its metabolites. Factors influencing the estimation of cocaine in sewage
364 with WWA have been evaluated. Results show that omitting in-pipe bio-
365 transformation affects the accuracy of back-calculated cocaine use estimates. In
366 addition, ex-vivo biotransformation of target compounds should be considered during
367 back calculation (Plósz et al., 2013). van Nuijs et al. evaluated the stability of nine
368 illicit drugs and metabolites in samples collected from wastewater influent. The
369 results suggest that it is quite important to take the compounds stability into account
370 when dealing with drugs that show significant biotransformation in sewage (van Nuijs
371 et al., 2012).

372 **3 Applicability in China**

373 **3.1 Research related to wastewater in China**

374 China has the largest population size (1.4 billion) in the world. The total sewage
375 created across the country is estimated as high as 280 billion litres per day
376 (calculation based on 200 L per capita per day), and most of the populated areas are

377 sewerage and connected to STPs. WWA could thus be used in the evaluation of illicit
378 drug consumption as well as alcohol, tobacco (Reid et al., 2011) and other chemicals
379 which are closely related to public health and social sustainability.

380 Recently a small number of WWA studies were conducted in China for estimating
381 illicit drug consumption. Lai et al. utilised WWA in Hong Kong in 2011 to evaluate
382 daily and diurnal patterns of illicit drug consumption in the megacity (Lai et al., 2013).
383 Khan et al applied WWA in mainland China for the first time to monitor the
384 consumption of 14 illicit drugs in 4 megacities with samples from 9 STPs covering
385 approximately 11.4 million inhabitants. The results demonstrate that China has
386 different drug consumption patterns to European countries. Even within China, the
387 difference in drug use between the north and south could be observed (Khan et al.,
388 2014). It is proposed that licit manufacture of drugs is more stringent and thus
389 distinguishing licit from illicit sources of drugs may be possible by further research on
390 isomer production ratios of the parent compounds and conducting chiral analysis of
391 wastewater samples. This could potentially lead to monitoring drug manufacture,
392 formulation, distribution and consumption (Daughton, 2011).

393 Several review articles (Liu et al., 2009; Liu, 2005; Nie et al., 2011; Xie et al., 2004;
394 Zhou et al., 2007) have presented a range of research on persistent organic pollutants
395 (POPs) and emerging pollutants such as pharmaceuticals and personal care products
396 (PPCPs) in the aquatic environment, particularly in regards to wastewater and how to
397 better manage these chemicals. The articles cover studies on the occurrence of POPs
398 and PPCPs in STPs (Fan et al., 2011; Zhao et al., 2011), pollutants removal
399 mechanism in STPs (Jiao et al., 2012), fate and degradation of certain group of POPs
400 or PPCPs particularly in regard to river water (Lian and Liu, 2013; Liu, 2011; Zhang,
401 2013), development and optimization of analytical method to qualitatively and
402 quantitatively determine pollutants in various matrices (Chen et al., 2011; Yuan et al.,
403 2013). However, most studies work ‘downstream’ focusing on environmental
404 outcomes rather than ‘upstream’ which could provide the ability to evaluate human
405 exposure and the associated health risks.

406 **3.2 Drug consumption and control in China**

407 Illicit drug abuse in China can be traced back to the 1760s during the Qing dynasty.
408 The number of drug users in China increased dramatically after the Opium War. Issue
409 of drug abuse re-emerged in the last two decades which is mainly attributed to global
410 drug trafficking activities during the implementation of governmental reform and the
411 open-door policies of the late 1980s (Lu et al., 2008; Qian et al., 2006). Evidence has
412 shown that over the past decade cocaine and other illicit drug abuse has increased in
413 East and Southeast Asia. The Ministry of Public Security in China estimates that there
414 are currently more than two million illicit drug users in China (Xinhua News, 2013).
415 Also, cocaine seizures in mainland China and Hong Kong increased from
416 approximately 600,000 kg in 2010 to 800,000 kg in 2011 (UNODC, 2013).
417 Illicit drug consumption has caused significant consequences for human health and
418 social stability. In response the Chinese government monitors illicit drug prevalence
419 and control. However, during the past decade the number of people abusing drugs has
420 increased significantly and younger generations have become victims of drug
421 addiction. Synthetic psychotropic drugs (like methamphetamine) prevail among drug
422 consumers. This situation requires the authorities to design effective policies to
423 control drug abuse as well as monitoring the effectiveness of these policies.
424 The ability of WWA to measure near real-time consumption of drugs can assist
425 authorities in assessing the impact of the strategies they've adopted and thus better
426 manage the situation. In order to develop and implement effective anti-drug strategies,
427 authorities need information about temporal and geographical patterns of illicit drug
428 consumption. Wastewater analysis could provide continuous and objective illicit drug
429 consumption information to the relevant authorities.

430 **3.3 Overview of sewage treatment plants and analytical laboratory capabilities in** 431 **China**

432 There are more than 3,000 domestic STPs in China covering most densely populated
433 areas. The number of STPs is still rapidly increasing with substantial investment from
434 the Chinese government to reduce environmental impacts. Capacity of these plants
435 range from less than 10 ML/day to more than 1,000 ML/day. Population size served
436 differs from a few thousand to hundreds of thousands. Most STPs have online

437 monitoring of flow, pH, COD and ammonia using auto-samplers and nearly all of
438 these plants take regular samples for compliance purposes which may make it easier
439 to get samples for WWA applications. Therefore, WWA could potentially capture
440 chemical consumption and/or exposure for a variety of population sizes with
441 considerably small effort and cost compared with traditional surveys.

442 China has strong analytical chemistry capabilities with hundreds of research centres and
443 laboratories located across the country equipped with state of the art analytical
444 instruments. There are more than 200 laboratories equipped with LC-MS/MS
445 instruments in various configurations and thus have sufficient analytical abilities to
446 apply WWA at the national level. Chen et al. (2011) developed a paper strip
447 extraction ultra-performance liquid chromatography tandem mass spectrometry (PSE-
448 UPLC-MS/MS) method to determine 9 PPCPs in sewage sludge. With further
449 optimization, this method could be suitable for WWA applications for drug
450 consumption estimation as well for the measurement of other chemical biomarkers of
451 consumption and exposure. More recently Yuan et al. (2013) developed and applied
452 an automated solid phase extraction-high performance liquid chromatography coupled
453 with electrospray ionization tandem mass spectrometry (ASPE-HPLC-ESI-MS/MS)
454 method for the quantification of 13 antipsychotics. Eleven of the thirteen
455 pharmaceuticals were detected in all 35 samples from one STP. Further studies on
456 wastewater treatment processes, human health biomarkers and risk assessment could
457 all benefit through promoting WWA as a feasible and powerful tool for forensic
458 science, environmental science and epidemiology.

459 **3.4 Potential issues with applying WWA/SCIM in China.**

460 There is no doubt that WWA can provide indicative information for the assessment of
461 illicit drug consumption. By sampling a variety of STPs and collaboration with the
462 many advanced research facilities across China, WWA/SCIM could produce valuable
463 information on current community health which could help define key areas of
464 concern for both community health and maintaining social justice. However,
465 investigation and assessment about the study area and objectives should be carried out
466 before conducting WWA to maximise results. Most STPs constructed before the
467 1990s receive influent that is a mixture of domestic sewage, industrial wastewater and
468 stormwater. This may make it more challenging to apply WWA in these areas as the

469 chemicals in the industrial sewage could interact with the chemicals in domestic
470 sewage and during rainfall events chemicals of interest may become too diluted to
471 analyse feasibly.

472 By comparing concentrations of target chemicals in ambient environmental
473 monitoring with results of available biomonitoring studies and WWA data, chemical
474 consumption/exposure models could be developed for pollutants chemicals,
475 biomarkers of human health, per capita environmental impact and others. One should
476 also consider that there are huge population relocations during certain national
477 holidays such as Chinese spring festival. Therefore real-time population estimates of
478 the studied catchment area is essential to reduce under/overestimation of the per
479 capita consumption and/or exposure of chemicals. These markers would also require
480 some form of calibration for the studied catchment such as collecting samples during
481 a census period.

482 As most of the STPs in a given city belong to a drainage group governed by the water
483 resource bureau or the environmental protection agency in the municipal government,
484 there might be concerns regarding ethical issues related to WWA studies. However, it
485 was suggested that WWA doesn't raise major ethical concerns when used for public
486 health purposes to monitor illicit drug use in large populations (Hall et al., 2012)
487 although ethical issues may arise from concerns about possible indirect harm from
488 using WWA in small areas such as prisons or entertainment venues. More effort is
489 required from the research community, industry and government departments to
490 promote WWA as an additional tool for illicit drug consumption monitoring.

491 **4 Conclusions**

492 Wastewater analysis is a promising approach to estimate illicit drug consumption and
493 consumption/exposure of other chemicals of concern at the population level. Our
494 review suggests that WWA could be a very useful tool in China. It could provide a
495 relatively easy approach for China to monitor drug consumption and potentially drug
496 trafficking and manufacturing. Early adoption of WWA/SCIM and archiving samples
497 would allow China to both make assessments using the current knowledge, as well as
498 create a sample bank that archives and allows reassessment of samples once analytical
499 methods are developed or new chemicals of interest are identified. Combined with

500 traditional survey methods, WWA could be a powerful tool to optimize illicit drug
501 consumption estimates and provide near real-time and objective data for the
502 development of strategies concerning drugs of abuse. With progress in research on
503 other WWA biomarkers, the approach will provide useful epidemiological data for
504 health status including levels of certain diseases in different communities and might
505 lead to the establishment of new monitoring approaches for population health.

506

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